

Carbon balance shift in mountain peatlands along a gradient of grazing disturbance in the tropical Andes (Colombia)

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Abstract

High-elevation cushion peatlands are typical ecosystems of the Andes above 4000 m of altitude, with an important role in hydrology and global carbon sequestration. In Tropical Andean context, grazing livestock is one of the main threats to cushion peatlands, altering the vegetation and the storage carbon function. The aim of this research is to understand how grazing influences cushion peatland functioning by identifying ecological thresholds for carbon balance process. The study was carried out in four Andean peatlands in the northern part of Colombian Andes during 2019-2020. We established 30 plots of 1 m² where water table level, vegetation cover, and grazing disturbance were monitored. We also measured CO₂ fluxes using an infrared gas analyzer connected to a closed static chamber, which registered net ecosystem exchange and respiration data. Considerable variation in the conservation status of *Distichia muscoides* cushions was found within the sampled peatlands, reflecting an heterogeneous signal of grazing disturbance that is evident at the plot-specific scale. Decreasing water table level was related with changes in dominant vegetation, from compact cushion species to grasses proliferation, exacerbating disturbance effects and carbon emissions. Mixed effects logistic regression models showed a carbon balance shift, from CO₂ sink to net emitter, in plots with high disturbance intensity and low *D. muscoides* cover. This study provides information for a better understanding of mountain peatlands functioning in the Tropical Andes and underlines the key role of *D. muscoides* cushions and the water table in carbon balance shift.

INTRODUCTION

Peatlands are common ecosystem in tropical mountains, and they have one of the largest carbon density of all the ecosystems, with nearly 1000 ton C ha⁻¹ (Hribljan et al. 2016). Peatland carbon storage requires the accumulation of partially decomposed organic matter over thousands of years, resulted from the imbalance between higher primary production and limited soil decomposition. Human disturbances such as grazing livestock affect the stability of the water table and plant composition, which are key elements that have a direct effect on the ability of the system to accumulate peat (Belyea and Malmer 2004; Moore and Dalva 2006). Our understanding on how grazing disturbance alter carbon accumulation process in the peatland ecosystems is still unknown in the tropical Andes. Better understanding of carbon balance shifts once it is pushed past a critical threshold is needed. Management strategies for Andean mountain peatlands relies on indicators that help address the impacts of grazing disturbance intensity for a better ecosystem management and conservation (Chimner and Welker 2011).

Carbon storage in the soil requires an organic matter input provided by the litter or old roots from peat-forming plants, such as *Distichia muscoides* which is peat-forming cushion plant with high net primary production rates but it's also dependent of flooding conditions (Cooper et al. 2015). Thus, changes in vegetation composition and water table depth results in the alteration of carbon assimilation and decomposition rates, meaning a net ecosystem uptake (Clymo 1984; Turetsky et al. 2002; Belyea and Malmer 2004). The alteration of the carbon balance happens when certain stress conditions cause greater respiration loss than the CO₂ uptake due to peat oxidation and acceleration of decomposition

rate. The change in the functioning of the system, from a carbon sink to a net emitter of CO₂ correspond to a cross of an ecological threshold (Turetsky et al. 2002; Moore and Dalva 2006; Larsen and Alp 2015; Toms and Villard 2015). Although, studies dedicated to understanding ecological thresholds from empirical data are limited and the scope in decision-making is questioned (Dodds et al. 2010; Hillebrand et al. 2020), this approach could be useful to represent critical conditions of carbon emissions and minimize adverse changes in the peat accumulation process by identifying which conditioning variables put the process in risk (Belyea and Malmer 2004; Johnson 2013; Toms and Villard 2015).

Even though mountain peatlands cover small areas, they have large carbon stocks in the soil due to a combination of high mineral content from the geomorphological process of the mountains and the type of vegetation that forms the peat (Hribljan et al. 2017). The steep topography of the Andes, together with a diverse and complex geomorphological setting of high-altitude lakes and glacial valleys, provide adequate conditions for the formation of wetlands and peatlands with unique vegetation (Cooper et al. 2015). Along the tropical Andes, at high elevations (> 4000 masl), peatlands vegetation grows in compact cushion dominated by vascular plants such as *D. muscoides* (Juncaceae) and *Plantago rigida* (Sklenář and Jørgensen 1999). Since cushion species at high elevations are the main peat-forming plant and play a key role in carbon sequestration, changes in cushions dominance could be catastrophic for the functioning of mountain peatlands.

Andean mountains and in particular their alpine ecosystems, are important sources of feed for native and non-native grazers, including Andean camelids at least for the last 9,000 years (Browman 1984), especially in dry Puna region of the central Andes, where peatlands provide the evergreen forage for the livestock in the dry season (Maldonado Fonkén 2014). Grazing plays an important role for biomass production because of the historical relationship between native camelids and cushion peatlands (Garcia et al. 2014). However, in the northern Andes, cushion peatlands are instead affected by introduced grazers such as sheep and cattle. Introduced grazers generate a greater disturbance on plant cover due to their heavier bodies and hard hooves (Maldonado Fonkén 2014; Machaca et al. 2018). Trampling by non-native consumers disaggregate and expose peat soil, removing part of the biomass and the carbon stored in peat due to aerobic decomposition (Salvador et al. 2015).

Here we hypothesized that human disturbances by high intensity grazing, combined with a lowering water table could negatively affect the carbon balance of cushion peatlands dominated by *D. muscoides*, shifting the system from a carbon sink to a carbon source. Therefore, the purpose of this study was to assess how grazing and vegetation influences CO₂ emissions in Tropical Mountain peatlands and the limit at which the system shift from a carbon sink to a net carbon source in order to provide management tools that allow preserving the processes of carbon accumulation of peatland ecosystem. Our research focused on 1) comparing environmental drivers and vegetation patterns along a grazing disturbance gradient, and 2) identifying response variables that affect carbon balance, using the ecological thresholds approach.

METHODS

Study site

The study was conducted at the high elevations of El Cocuy National Park, in the northern Andes of Colombia (Fig. 1). A large part of the protected area is overlapped with Indigenous territory, as well as private lands managed by farming families who frequently grazed their livestock within the protected area. There is an annual mean temperature of 6.4 °C and a bimodal precipitation regime, with two wet periods in May and October with an average rainfall of 868,4 mm (López-Moreno et al. 2022). We randomly selected four peatlands dominated by cushions of *D. muscoides*, combined with graminoids such as *Agrostis breviculmis* and *Calamagrostis effuse* and the moss *Breutelia chrysea* (Fig. 2). All sites had more than five meters of peat and were located at a similar elevation and a long history of livestock use, with evidence of sheep grazing disturbance (Table 1). The peatlands were visited on five occasions during the 2019–2020 period to include the annual climatic variation.

Table 1
Descriptions of the studied cushion dominated peatlands sampled at El Cocuy National Park. Geographical coordinates, altitude and number monitored plots

Sites	Geographical coordinates	Altitude (m)	Num. plots
Chapetona	6°29'39.44"N 72°21'03.77"W	4200	5
Curial	6°31'38.13"N 72°20'57.23"W	4350	8
Curial Alto	6°31'55.34"N 72°20'43.57"W	4450	8
Corral Grande	6°32'50.41"N 72°21'20.02"W	4450	10

Vegetation sampling and environmental factors

We located a total of 30 1 m² plots randomly positioned and stratified across the main plant and ground cover gradients observed on the selected peatlands, the cover was classified in: cushions of *D. muscoides*, hollows, dead cushions, grass dominated sites and moss dominated sites. Plant composition was sampled in all the plots using a 1 m² quadrant with 10 x 10 cm sub-quadrants (Mueller-Dombois and Ellenberg 1974), where plant and ground cover was estimated to the nearest 1% for each species. Overall, the percent cover of *D. muscoides*, moss, grasses, bare ground, hollows, standing water, and exposed peat was estimated. Plant species were identified following standard herbarium identification protocols, and species names were assigned following the international plant name index (IPNI 2023. <https://www.ipni.org/> accessed February 28, 2023).

Due to the differential sheep grazing intensity within a single peatland, disturbance was not homogenous throughout the site. Thus, disturbance was characterized at plot level and nearby area using the percentage of exposed peat cover and the evidence of sheep alteration such as excrements presence, hoof marks, trails and/or herbivory. The type of specific disturbance was classified according to a semi quantitative intensity scale from 1 to 6, being the presence of excrement the lowest intensity and, the

most severe, disaggregated exposed soil (Online Resource: FigS1). The disturbance category was developed from combining exposed peat cover and the sheep alteration scale. For water table level we installed 30 monitoring wells that were located near the 1m² plots. Water level was measured manually during the different visits to the study site. When the water table was above the peatland surface a positive value was recorded and, if it was below the surface, a negative value was recorded. The water table level was monitored for each plot at each visit to the study site. Also, in each visit a water sample was collected to measure pH and electrical conductivity (EC). Average monthly air temperature and precipitation were provided by the Instituto de Hidrología, Meteorología y Estudios Ambientales (IDEAM) from a meteorological station closest to the study sites.

CO₂ Measurements

CO₂ flux was measured using an infrared gas analyzer (EGM4; PP Systems, Amesbury, USA) connected to a closed transparent static chamber of 60 cm of diameter and a height of 50 cm (Hutchinson and Mosier, 1981). For each measurement, soil and air temperature, relative humidity, atmospheric pressure, and photosynthetically active radiation using a PAR sensor attached to the inside of the chamber were measured (Apogee MQ-200 Quantum meter). Using a. A total of 105 flux measurements with a duration of 124 seconds each were performed in two consecutive phases: 1) a phase where the chamber receives natural solar radiation, registering net exchange of the ecosystem (NEE), and a second phase where the chamber does not receive radiation, registering autotrophic and heterotrophic respiration (RE). Gross primary production (GPP) was calculated as the difference between net ecosystem exchange and ecosystem respiration for each pair of light and dark, flux measurements.

Statistical analysis

Parametric and non-parametric analyzes of variance (one way-ANOVA and Kruskal-Wallis) were performed to identify differences between the environmental and flux variables among the sampled peatlands. Plant composition was analyzed using a cluster analysis with UPGMA method as well as analysis of similarity between sites and vegetation plots based on the Bray Curtis dissimilarity matrix among samples. Additional MDS analysis was used to analyze the homogeneity differences in plant composition and the main environmental gradients associated to the changes in species composition. The main environmental patterns were analyzed using principal Components Analysis (PCA). A repeated measures analysis with mixed effects was applied using lme4 package to quantify the relationships between explanatory variables (vegetation cover and water table) and CO₂ fluxes, where plot and period were incorporated in the models as random effects, and maximum likelihood estimation was performed. CO₂ fluxes was the dependent variables with GPP, NEE and ER analyzed separately. A few NEE outliers were omitted due to equipment problems.

The relationships between vegetation and water table and NEE were analyzed using mixed effects binomial logistic regression models to identify ecological thresholds (McCullagh and Nelder 1989; Toms and Villard 2015). NEE was coded as a binary variable, converting negative values to values of 0 (net ecosystem uptake) and positive values to values of 1 (net ecosystem emission). The threshold was

determined when the probability of the system going from sink to emitter crossed 0.5 probability (Toms and Villard 2015). The selection of the logistic models that better explained the changes in NEE was made based on the significance of the model and AIC value. Residuals were checked for normality and that homogeneity of variance assumptions were met, and significance was determined at an alpha value of 0.05. All data analyses were performed with the open-source software package R, version 4.1.3 2022.

RESULTS

Vegetation patterns and environmental factors

A total of 43 species were registered, including 13 bryophytes and 30 vascular plants. Plot vegetation was dominated by *D. muscoides* with a mean 54% cover in all study sites ($F_{4,130} = 35.2$, $p = 0.001$; Table 2). Exposed peat was more common in Curial alto and Corral Grande (15 and 16% cover, respectively), and commonly associated with sedges (*Carex* sp.), grasses (*Agrostis breviculmis*), rosettes such as *Werneria pygmaea*, herbs that were commonly on disturbed non-peatland areas such as *Lachemilla nivalis* and *Oreomyrrhis andicola*, and mosses such as *Bryum argenteum* and *Breutelia* sps. According to obtained data, five cover types were classified in the plots: 1) *D. muscoides*, 2) mosses, 3) grasses, 4) exposed peat, and 5) other vascular plants (Table 2).

Table 2 Summary of main values and SE of hydrology, vegetation cover, soil carbon content and carbon gas flux data from the four study sites. Soil carbon content is expressed in Mg C ha^{-1} whereas net ecosystem exchange (*NEE*), ecosystem respiration (ER) and gross primary production (GPP) are expressed in $\text{CO}_2 \text{ g m}^{-2} \text{ h}^{-1}$

	Chapetona	Curial	Curial alto	Corral grande
Hydrology				
Water table (cm)	-15.3.0 ± 8.9	-9.7 ± 7.1	-17.8 ± 11.1	-15.1 ± 11.8
pH	6.4 ± 0.3	6.84 ± 0.1	6.6 ± 0.1	6.3 ± 0.1
EC (μS)	64.8 ± 38.8	84.0 ± 36.7	71.1 ± 31.1	30.9 ± 8.4
Vegetation				
Number of species	28	24	25	34
<i>D. muscoides</i> cover (%)	52.8 ± 22.6.0	60.1 ± 20.7	54.2 ± 20.9	49.6 ± 15.6
Moss cover (%)	25.0 ± 13.4	10.8 ± 8.9	7.5 ± 3.3	15.2 ± 8.1
Grass cover (%)	5.5 ± 3.5	11.1 ± 7.3	9.3 ± 7.2	5.8 ± 2.4
Exposed peat cover (%)	0.5 ± 0.9	6.8 ± 6.2	14.7 ± 11.7	16.3 ± 11.7
Other vascular plants (%)	16,2 ± 3.1	11,2 ± 12.2	14,3 ± 7.5	13.1 ± 4.
Peat properties				
Bulk density (g m ⁻³)	0.11 ± 0.03	0.06 ± 0.02	0.15 ± 0.03	0.09 ± 0.01
Carbon percentage (%)	43	36	43	45
Carbon content (Mg C ha ⁻¹)	121.7 ± 73.7	56.8 ± 19.7	106.4 ± 32.9	62.5 ± 10.2
CO₂ fluxes				
NEE (g CO ₂ m ⁻² h ⁻¹)	-0.08 ± 0.04	-0.01 ± 0.06	-0.001 ± 0.04	-0.06 ± 0.02
RE (g CO ₂ m ⁻² h ⁻¹)	0.44 ± 0.1	0.31 ± 0.2	0.37 ± 0.1	0.37 ± 0.1
GPP (g CO ₂ m ⁻² h ⁻¹)	-0.50 ± 0.2	-0.32 ± 0.1	-0.37 ± 0.1	-0.42 ± 0.2

Water table monitoring showed that site and time of the year had a significant effect on water table depth, being Curial site the wettest, followed by Corral Grande, Chapetona and Curial Alto the driest. Water table levels ranged from 10 to 33 cm below the soil surface, and were very variable throughout the year, with a mean depth of -19.6 ± 12.2 cm during the wet season, and -30.6 ± 10.3 cm during the dry season (January and February), when some piezometers were completely dry (Table 2). Plots with less exposed peat showed more stable water table levels throughout the year ($F_{4,130} = 35.2$, $p = 0.001$).

Water table showed a significant positive relationship with *D. muscoides* cover from the mixed model ($Pseudo-R^2 = 0.5$, $p = 0.005$), and a negative relationship with grasses ($Pseudo-R^2 = 0.2$, $p = 0.06$). Plots with water table above 20 cm depth had up to 50% *D. muscoides* cover, whereas plots where the water table was lower (< 20 cm from the surface), had an increase of 15% cover of grasses (Fig. 3). A negatives relationship between water table depth and the percentage of exposed peat was also observed ($Pseudo-R^2 = 0.38$, $p = 0.03$), except in Chapetona site, where almost no exposed peat was recorded.

The ordination diagram shows there was a plot-scale heterogeneity regardless of the peatland. Plant composition changed across the water table and disturbance gradients. Plots with higher exposed peat

cover had higher dominance of species associated with drier conditions such as *Agrostis breviculmis*, *Lachemilla nivalis*, *Bryum argenteum*, and *Oreomyrrhis andicola*. In contrast, plots with higher water table level and low disturbance intensity were dominated by *D. muscoides* and *Plantago rigida* and had less exposed peat cover (Fig. 4). Exposed peat cover was strongly correlated with disturbance categories ($r = 0.89$) and was the environmental factor that better explained variance between plots.

Carbon fluxes

The studied peatlands had an average carbon assimilation rate of $-0.044 \pm 0.17 \text{ g CO}_2 \text{ m}^{-2} \text{ h}^{-1}$. Carbon fluxes were positively related to soil temperature, solar radiation, and water table level. GPP ranged from -0.16 to $1.74 \text{ g CO}_2 \text{ m}^{-2} \text{ h}^{-1}$ (Table 1). GPP was influenced by PAR, with a significant response function during the hours of highest radiation (around $1900 \mu\text{mol m}^{-2} \text{ s}^{-1}$; Fig. 5). Carbon assimilation values decrease at low water table levels and increase with increasing water table level ($Pseudo-R^2 = 0.29$, $p = 0.04$; Fig. 5). In contrast, ER responded oppositely to increasing water tables ($Pseudo-R^2 = 0.45$, $p = 0.5$). Also, ER had a strong relationship with soil temperature, since ER values decrease with soil warming ($Pseudo-R^2 = 0.35$, $p = 0.12$; Fig. 5).

D. muscoides, exposed peat and grass cover had a significant relationship with NEE. Carbon assimilation decrease in exposed peat-dominated plots, where positive or low NEE values were recorded ($Pseudo-R^2 = 0.13$, $p = 0.01$). In contrast, plots with high *D. muscoides* cover had negative NEE values, which means high carbon assimilation rate ($Pseudo-R^2 = 0.15$, $p = 0.0001$). Also, high cover of grass was related with positive NEE values. The logistic regression models showed a variation on carbon balance values related to the loss of vegetation cover of the cushions at the plot scale. Both when *D. muscoides* cover decrease and when the exposed peat area increases, a threshold point was found for which there was a greater probability that respiration exceeded assimilation. Plots with less than 33% of *D. muscoides* cover, had a high probability of a net carbon emission, reaching a 0.78 probability of being net carbon emitters when *D. muscoides* cover was completely lost (Fig. 6). In the same way, plots with more than 25% cover of exposed peat had also a high probability of net carbon emission (Fig. 7).

DISCUSSION

The variation in species composition along the disturbance gradient suggests that the signal for grazing disturbance is heterogeneous within the peatlands, and it is evident at plot-scale rather than on the whole peatland as a unit. The relationship between low water table and high grass cover also could explain the replacement of cushion species: graminoids need more drained conditions to proliferate, which could be a direct effect of trampling and grazing, reducing cushions surface humidity (Machaca et al. 2018; Zhong et al. 2020). Plants composition changes and water table fluctuation could lead different trajectories of plants composition, by suppressing cushions species and favoring the proliferation of species more adapted to grazing use and drier condition such as *Agrostis breviculmis*, *Lachemilla sps* and *Oreorrhymys andicola* (Benavides 2014; Loza Herrera et al. 2015; Greenwood et al. 2019). Also, oral communications with local people and field observations showed that sheep did not forage *Distichia sp.*

itself, but rather the graminoids growing over the cushions. This interaction could lead to positive feedback between graminoids proliferation favored by water table drawdown and increasing disturbance, opening new questions such as whether the presence of grasses is explained, to a greater extent, by disturbance or by the low water table.

Our results for net ecosystem uptake were in a range of -0.001 and $-0.67 \text{ g CO}_2 \text{ m}^{-2} \text{ h}^{-1}$, much lower than those reported for cushion-dominated grazed peatlands in Ecuador ($-1.25 \text{ g CO}_2 \text{ m}^{-2} \text{ h}^{-1}$) (Sánchez Garces et al. 2017). Besides both studies being carried out at similar altitudes ($> 4200 \text{ m}$) and hydrological conditions, the Ecuadorian peatland barely had exposed peat cover as well as non-cushion plant abundance, which could explain the carbon balance differences (Urbina and Benavides 2015; Sánchez Garces et al. 2017). Water table also showed a spatial relationship with carbon balance values, with high ER values and low positive NEE values (lower C uptake) at low water table level (Zhong et al. 2020). Water table drawdown increases the oxygen diffusion within the unsaturated peat and promotes peat oxidation by an increase in soil microorganism ER activity, producing carbon loss (Moore and Dalva 2006; Jaatinen et al. 2008; Hribljan et al. 2014).

This research allow us to understand the influence of water table fluctuations on the instantaneous CO_2 fluxes, highlighting the importance of short-term variation understanding to infer the effects of water table drawdown if it is more prolonged in time. Besides the explored methods were limited to explain the variability of peat accumulation process at large temporal scales, drier conditions could increase decomposition rate of peat and grasses proliferation, leading to a rapid loss of stored carbon and driving to an ecosystem shift from carbon sink to a net carbon emitter. Although cloud cover was not considered in the analysis, it has a significant effect on carbon balance shift, having a positive impact on CO_2 assimilation rate (Planas et al. 2020). However, the one-year monitoring of CO_2 fluxes expect to lead us to include annual climatic variation and rule out the effect of cloudiness over the NEE results.

With one-year observations, the logistic regression models showed a carbon unbalance related to the disturbance signal at a plot scale, from which respiration exceeded assimilation. Carbon balance shift from negative to positive *NEE* was observed both for the decrease in *D. muscoides* cover, and the increase of exposed peat cover, pointing out an ecological threshold from C sink to C source (Pelletier et al. 2015). Areas with less than 33% of *D. muscoides* cover and more than 25% of exposed peat cover are likely to have crossed the ecological threshold and they are losing the stored carbon, compromising the persistence of the ecosystem (Sánchez Garces et al. 2017)

For ecological thresholds models, single drivers were analyzed following the best statistical models, confounding variation for the identification of threshold responses (Simmonds et al. 2017). However other abiotic variables could interfere, highlighting the importance of considering complementary drivers that also affect or interact with *NEE* such as water table level, microtopography, water chemistry and climate variability (Belyea and Malmer 2004). Furthermore, grazing pressure combined with environmental changes could bring forward the carbon balance shift at lower disturbance intensity

(Machaca et al. 2018), which could lead us to devise management strategies that delay threshold points such as develop cushion restoration and restrict livestock access at drier season.

Little is known about how ecological processes response to threshold cross or the capacity of vegetation and humidity to recover after certain disturbance intensity (Benavides et al. 2013). Evidence of disturbance effect will be dependent on the spatial scale measure, where the response variable and the environmental drivers selected will determine the identification of a threshold for carbon balance shift (Belyea and Malmer 2004). Despite better understanding of the resilience of the systems is required, small spatial and temporal scale processes such as *NEE* could be an early-warning signals for ecosystem decline (Eppinga et al. 2009).

CONCLUSION

This study provides information for a better understanding of the functioning of mountain peatlands of the Tropical Andes, crucial in CO₂ storage and water regulation (Salvador et al., 2014), and underlines the key role of *D. muscoides* cushions and the water table in the peat accumulation process. Considerable variation in the conservation status of *D. muscoides* cushions was found within the sampled peatlands, showing an heterogenous signal of grazing disturbance and an effect on the carbon balance at microscale level. Since carbon balance determines a key function of peatlands, high-resolution variables like *NEE* could be a useful response variable to anticipate responses in larger-scale processes. (Eppinga et al., 2007). This hypothesis needs further exploration through additional data for a better characterization of spatial patterns on carbon balance behavior. Rather than providing specific number of vegetation cover at which the system cross non-return state, these results could guide decision makers about consistent monitoring of specific response variables, and appropriate spatial and temporal scale at which thresholds may arise (Huggett, 2005).

Declarations

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Data Availability Statement

The data presented in this study are available on request from the corresponding author. The species data are attached in supplementary material.

Competing interests

The authors declare that the research was conducted in the absence of any commercial or financial interest to disclose.

Author contributions

Study conception and methodology: VMA and JC. Research design, data collection and samples processing: VMA. Formal analysis and data curation: VMA and JCB. Writing-original draft preparation: VMA. Writing—review and supervision: JCB. Project administration and funding acquisition: VMA and JCB. All authors reviewed the manuscript and approved the submitted version.

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Figures

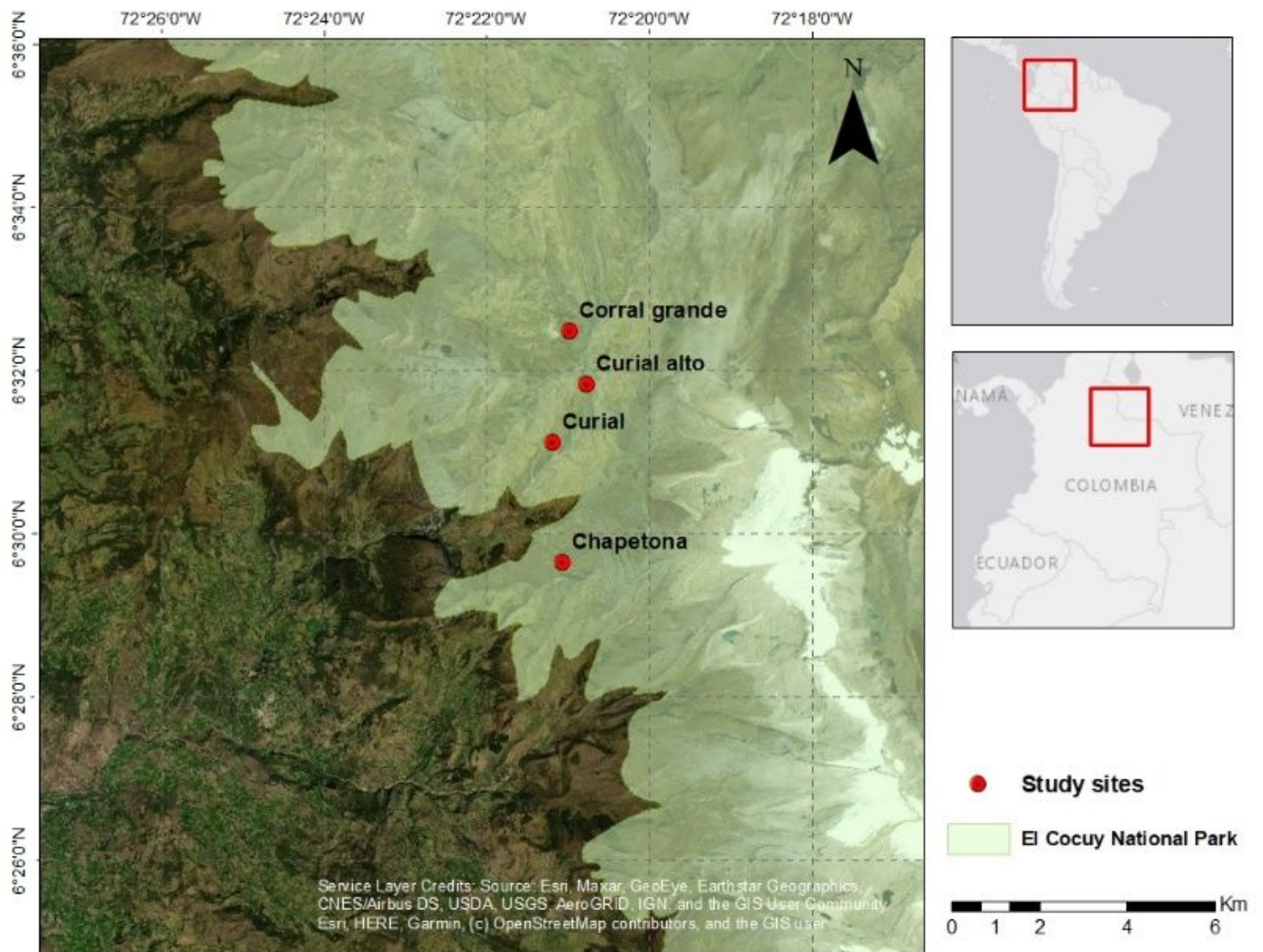


Figure 1

Location of study sites in El Cocuy National Park, in the northern Andes of Colombia

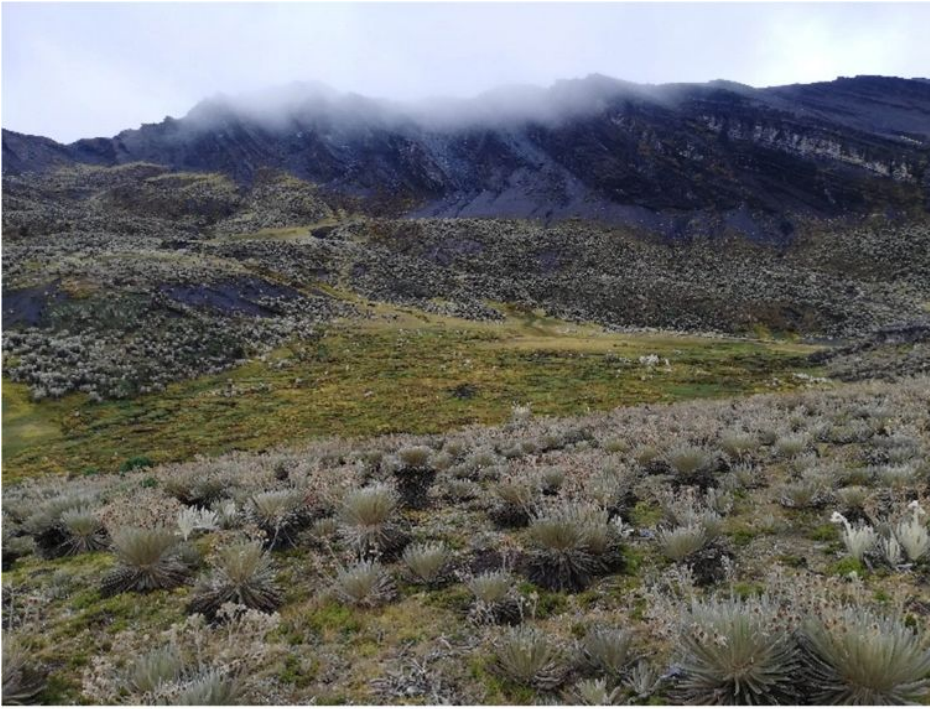


Figure 2

Cushion peatlands of *D. muscoides* in the sampled sites, located in El Cocuy National Park, northern Andes of Colombia. (Top left) Chapetona; (top right): Corral Grande; (bottom left): Curial and (bottom right): Curial alto

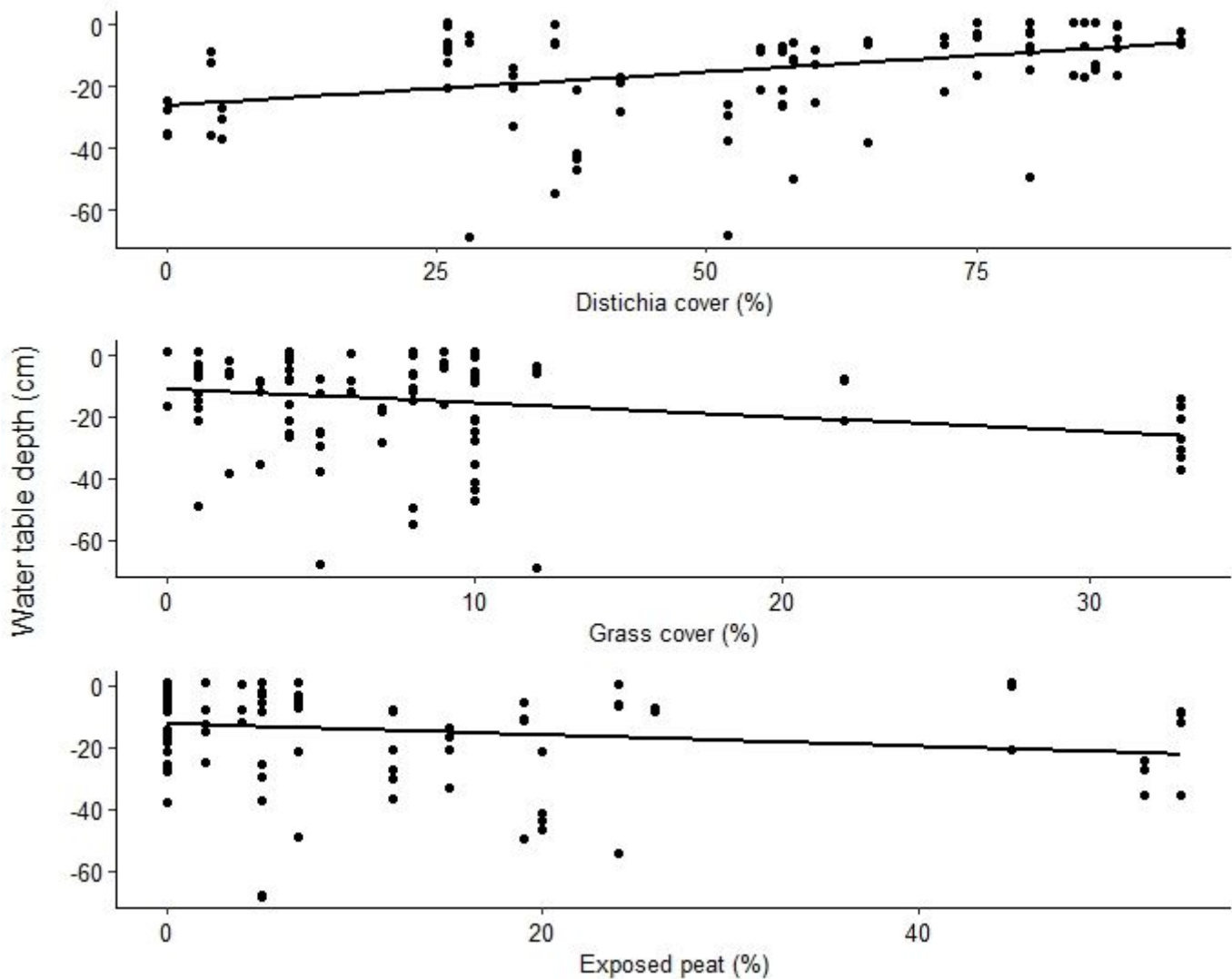


Figure 3

Linear relationship between water table level (cm) and vegetation cover type (%) for all sampled peatlands of the Cocuy National Park. (Top): *D. muscoides* cover; (center): grass cover and (bottom): exposed peat cover

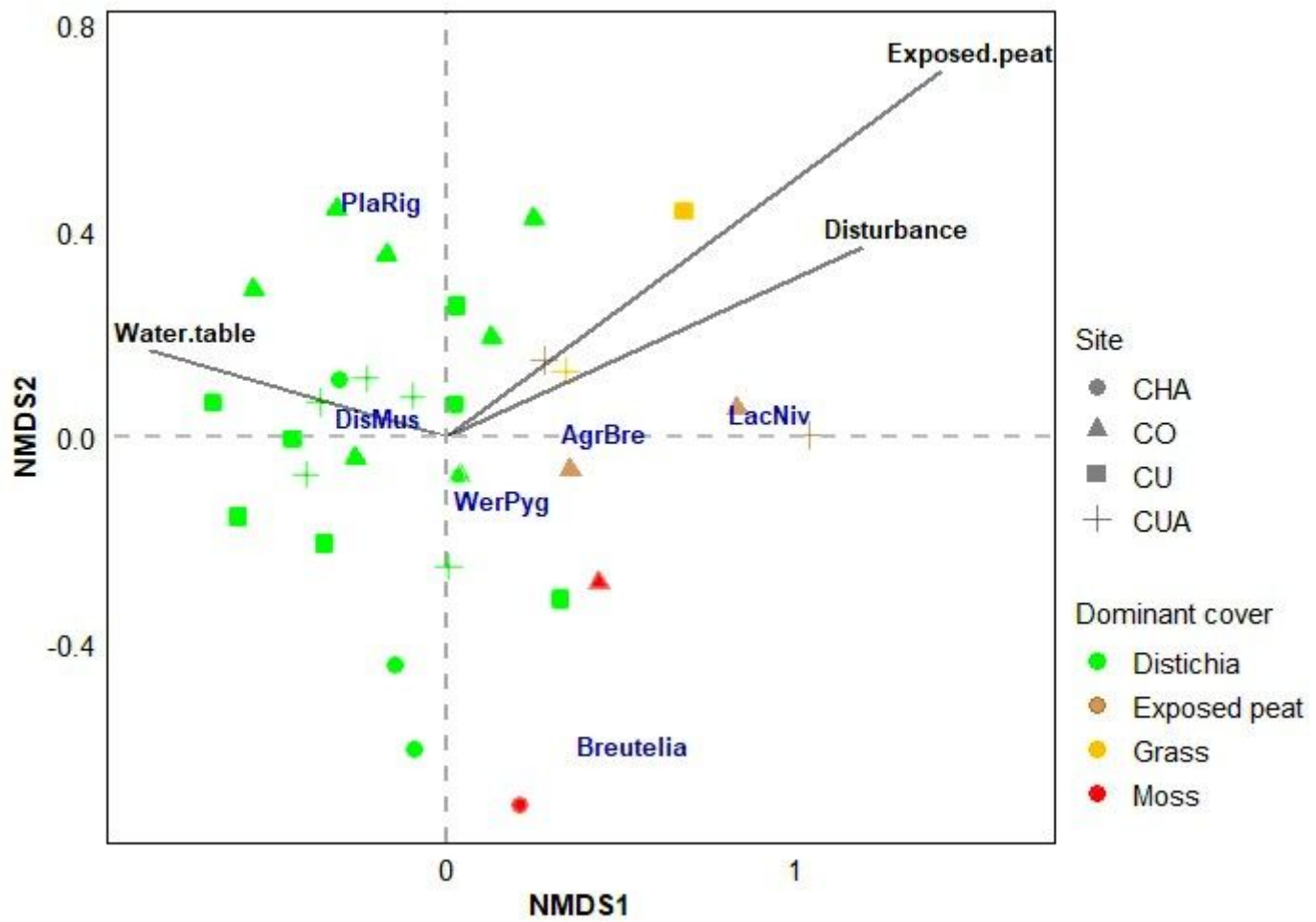


Figure 4

NMDS Analysis for the four sites of the Cocuy National Park: Chapetona (CHA), Curial (CU), Curial grande (CUA) and Corral Grande (CO)

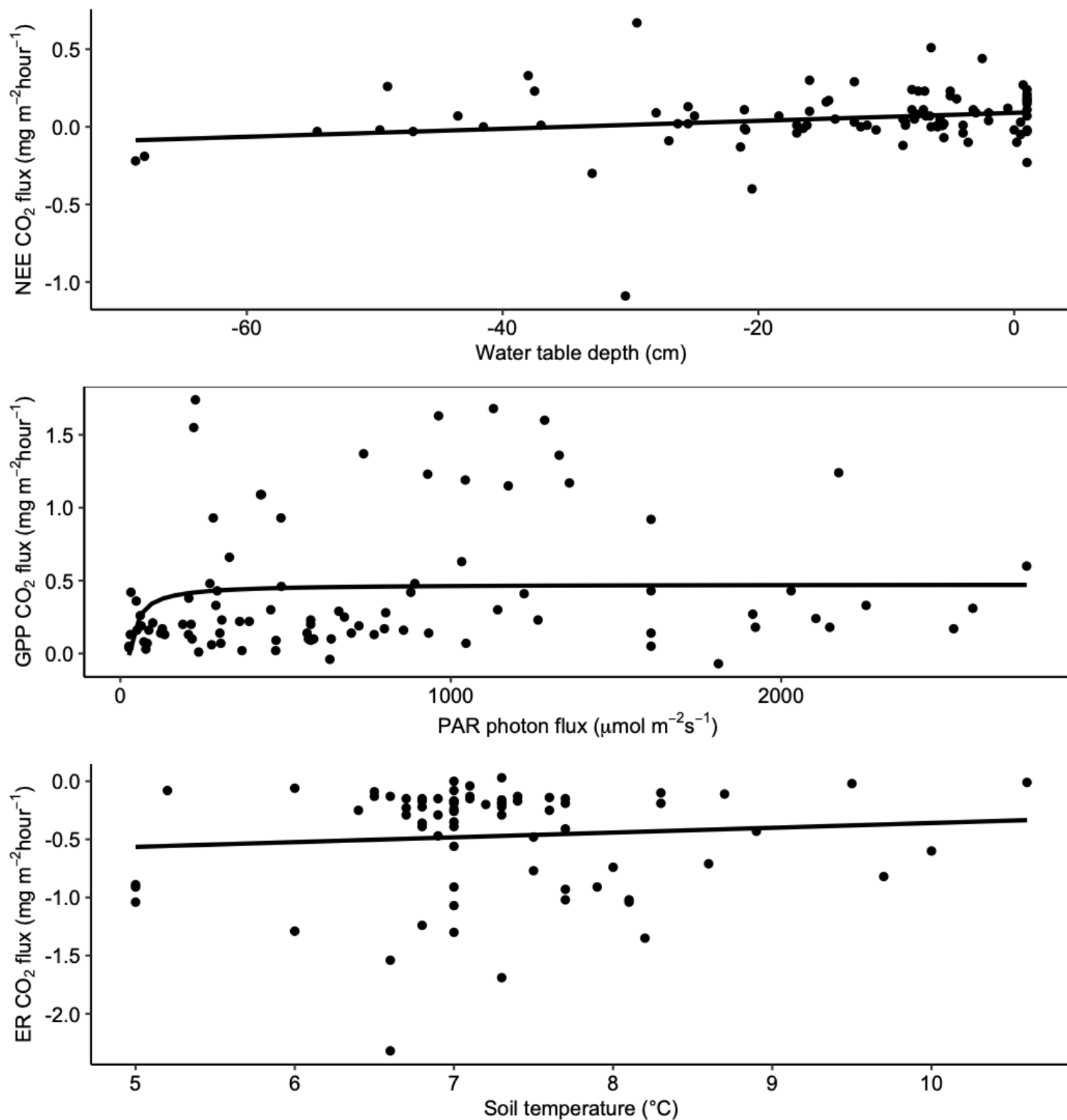


Figure 5

Linear relationship between net ecosystem exchange (NEE; top) and water table level (cm); gross primary production (GPP; center) versus photosynthetically active radiation (PAR; $\mu\text{mol m}^{-2} \text{s}^{-1}$); and ecosystem respiration (ER; bottom) and soil temperature ($^{\circ}\text{C}$) for all sampled peatlands of the Cocuy National Park

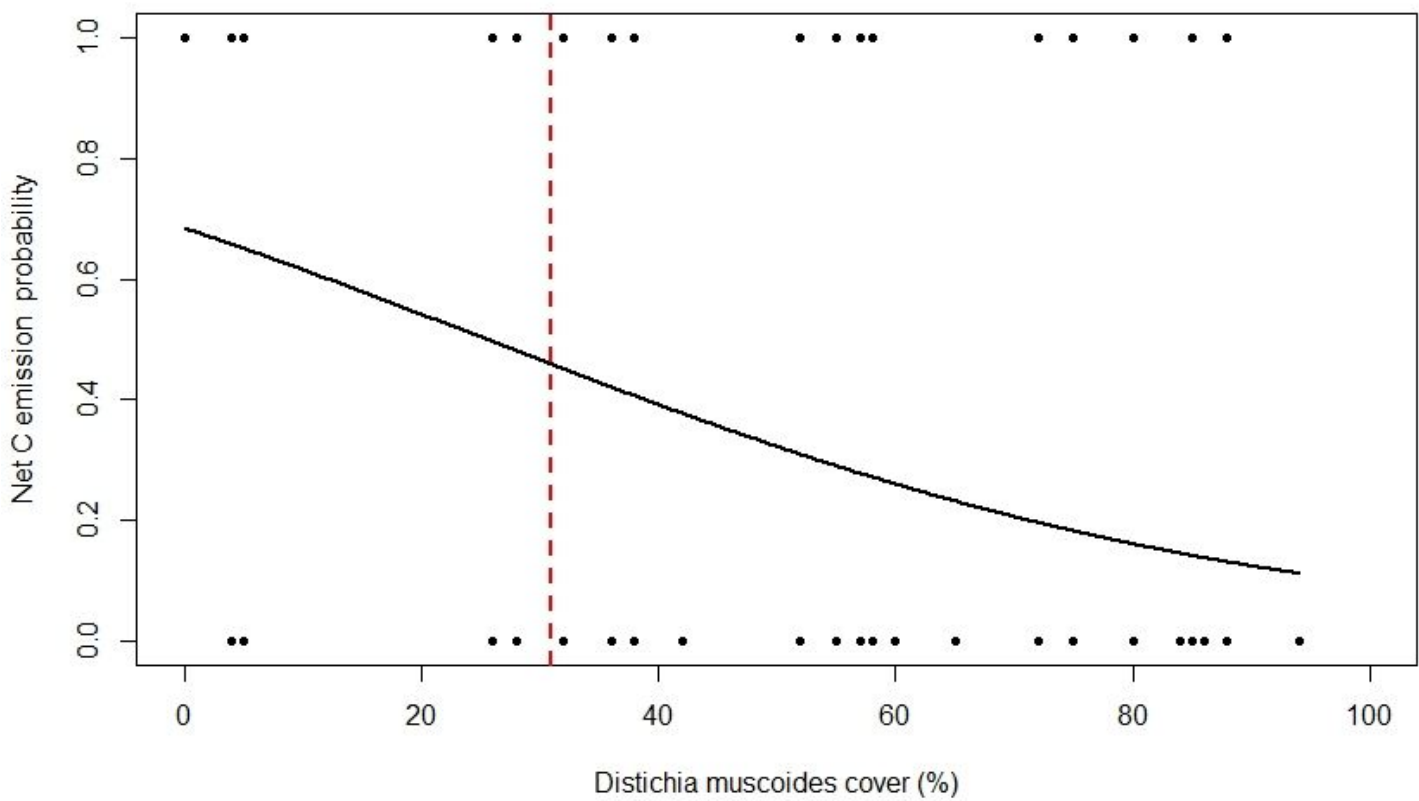


Figure 6

Probability of net carbon emission according to *D.muscoides* cover. Red line represents the ecological threshold where there is a 50% probability of a positive carbon balance for all sampled peatlands of the Cocuy National Park

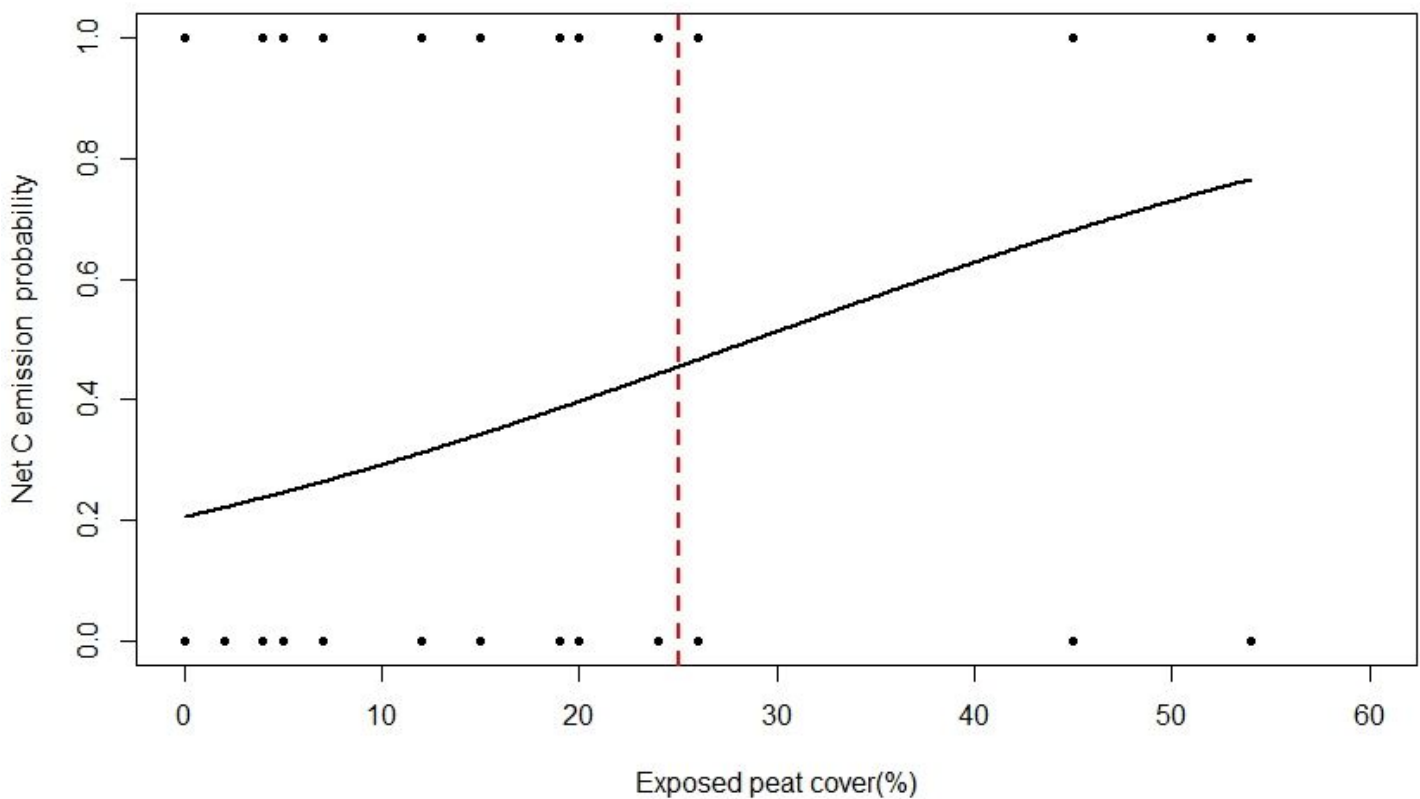


Figure 7

Probability of net carbon emission according to disturbance categories. Red line represents the ecological threshold where there is a 50% probability of a positive carbon balance for all sampled peatlands of the Cocuy National Park

Supplementary Files

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